Seasonal changes in the PM₁ chemical composition north of Mexico City

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RESUMEN

Se instaló un Monitor de Especiación Química de Aerosoles (ACSM, por sus siglas en inglés; Aerodyne Inc.) en un sitio al norte de la Ciudad de México del 13 de noviembre de 2013 al 30 de abril de 2014, con el objetivo de investigar la variabilidad estacional de la composición química del aerosol menor a 1 µm. El ACSM determina la concentración en masa, en tiempo real, de las especies más importantes (nitrato, sulfato, amonio, cloruro y compuestos orgánicos) del material particulado no refractario menor a 1 µm (NR-PM₁), con una resolución temporal de 30 min. Durante dicho periodo también se midieron variables meteorológicas (temperatura, humedad relativa, y dirección y velocidad del viento), y las concentraciones de carbono negro, PM₁, PM_{2.5}, CO, SO₂, NO, NO₂ y O₃. La concentración en masa de NR-PM₁ sumada a la concentración del carbono negro (que debe ser cercana a la concentración total de PM1) tuvo una buena correlación con la concentración medida con un equipo de microbalanza de elemento oscilatorio, lo cual indica la buena calidad de los datos del ACSM. En promedio, la composición del aerosol, así como su variabilidad diurna fueron similares a lo encontrado en campañas anteriores en las que se usó instrumentación similar (MCMA-2003 y MILAGRO). Sin embargo, el aerosol mostró un carácter ácido durante noviembre y diciembre, probablemente debido a una mayor humedad relativa, menor temperatura, y vientos más frecuentes del noroeste (donde se encuentra el complejo industrial Tula) durante este periodo. Una baja concentración de amoniaco en la fase gas (NH₃) también puede tener un efecto importante en la acidez observada. Estos resultados sugieren un cambio estacional en la química del aerosol, el cual debe verificarse llevando a cabo más estudios a largo plazo.

ABSTRACT

An Aerodyne Aerosol Chemical Speciation Monitor (ACSM; Aerodyne Inc.) was deployed at a site north of Mexico City from November 13, 2013 to April 30, 2014 to investigate the seasonal variability of the chemical composition of submicron particles. The ACSM provides real time information on mass concentration of the non-refractory main species (nitrate, sulfate, ammonium, chloride and organic compounds) in particulate matter less than 1 µm (NR-PM₁) with a 30-min time resolution. Meteorological variables (temperature, relative humidity, and wind speed and direction), as well as concentrations of black carbon, PM₁, PM_{2.5}, CO, SO₂, NO, NO₂, and O₃, were also measured. The total NR-PM₁ mass concentrations plus black carbon (which must be close to the PM₁ total mass) showed a good correlation with PM₁ mass concentration measured with a Tapered Element Oscillating Micro-Balance, an indication of the soundness of the ACSM data. In average, the composition of the aerosol as well as its diurnal variability were similar to observations in previous studies using

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similar instruments (MCMA-2003 and MILAGRO). However, it was observed that the aerosol was persistently acidic during November and December probably due to a higher relative humidity, lower temperature, and more frequent winds from the NW, where the Tula industrial complex is located. A lower concentration of ammonia (NH₃) in the gas phase might affect the PM acidity too. These results suggest a seasonal variability in the aerosol chemistry in Mexico City, which should be verified with more long-term studies.

Keywords: ACSM, NR-PM₁ chemical composition, aerosol acidity, Mexico City.

1. Introduction

The Mexico City Metropolitan Area (MCMA) is on average 2240 masl at 19° 29' N and 99° 90' W, in a basin surrounded by mountains, but relatively open to the north. There are more than 21 million inhabitants, and every year are emitted close to 27 706 ton of PM_{10} and 9 847 ton of $PM_{2.5}$ to the atmosphere (SEDEMA, 2016a). Being one of the largest megacities in the world, the MCMA can have significant impacts on human and ecosystems health, as well as on global change (Baklanov et al., 2016). Because of this, two major campaigns in the city have been organized focusing on understanding the processes and global effects of the city emissions: MCMA-2003 (Molina et al., 2007) and MILAGRO (Molina et al., 2010). MCMA-2003 was carried out in the spring of 2003 and, among other instruments, an Aerodyne Aerosol Mass Spectrometer (AMS) was deployed for the first time in Mexico with the objective of studying the chemical composition of the aerosol with a high time resolution. In such study, Salcedo et al. (2006) found that the organic fraction of the Mexico City aerosol was the largest, followed by the inorganic fraction (nitrate, sulfate, ammonium, and chloride). In general, sulfate and nitrate were neutralized by ammonium, except for some short periods with high sulfate concentration, which showed an acidic nature. The diurnal variability of the aerosol components was also analyzed showing that nitrate concentration was dependent on the photochemical activity, while sulfate concentration had a regional character. The variability shown by the organic fraction was explained by a combination of primary emissions and secondary production. During MILAGRO, in the spring of 2006, Aiken et al. (2009) again deployed an aerosol mass spectrometer in Mexico City. Their observations were consistent with those from MCMA-2003. In addition, they performed a detailed analysis of the organic fraction of the aerosol, identifying four components: a primary hydrocarbon-like component (HOA); an oxygenated component, probably of secondary origin (OOA); a component corresponding to biomass burning (BBOA); and a small local nitrogen-containing fraction (LOA). Both campaigns took place during the spring, for only few weeks, in one specific site each one. Since 2006, there has not been another similar study. Hence, on spite of all the knowledge generated during MCMA-2003 and MILAGRO, there is still a lack of information regarding the seasonal variability of the aerosol chemical properties in Mexico City.

In this paper, we present the results from groundbased measurements at a site north of Mexico City (very close to the MILAGRO supersite, T0) for a period of almost six months, during the winter of 2013-2014. PM₁ chemical concentrations and composition were measured with an Aerosol Chemical Speciation Monitor (ACSM), the most robust version of the AMS. This data, together with criteria pollutants concentrations and meteorological variables, was analyzed in order to look for seasonal changes in the aerosol chemistry. The ACSM data was compared to results from Salcedo et al. (2006) and Aiken et al. (2009) from MCMA-2003 and MILAGRO campaigns, respectively.

2. Methodology

2.1 Site

Data was collected from November 13, 2013 to April 30, 2014, at the Laboratorio de Análisis Ambiental (LAA) of the Secretaría del Medio Ambiente de la Ciudad de México (Environmental Analysis Laboratory of Mexico City's Secretariat for the Environment), north of Mexico City (19° 29' 01.19" N, 99° 08'5 0.08" W, 2255 masl). LAA is located within a mixed neighborhood surrounded by residential, industrial, and commercial areas; a road with heavy traffic is 200 m away from it. Figure 1 shows the location of the LAA site as well as the MCMA-2003 and MILA-GRO super-sites (CEN and T0, respectively). LAA and T0 are approximately 600 m apart.



Fig. 1. Map of the Mexico City Metropolitan Area showing the location of the LAA, T0, and CEN sites, as well as the Tula industrial complex. Black lines mark municipal and state boundaries; urban areas are colored in dark grey (INEGI, 2014).

2.2 Aerosol Chemical Speciation Monitor

The Aerosol Chemical Speciation Monitor (ACSM; Aerodyne Research Inc., Billerica MA) is an aerosol mass spectrometer (AMS) capable of measuring chemical composition of non-refractory submicron particulate matter (NR-PM₁) (Ng et al., 2011), which includes all species that can evaporate in few seconds at ~600 °C under vacuum (the temperature of the vaporizer within the instrument). Specifically, the ACSM detects nitrate, sulfate, ammonium, chloride, and the organic fraction of the aerosol; it does not detect components such as sea salt, soil dust, and elemental carbon (Jiménez et al., 2003). For this study, the ACSM was calibrated following the procedures described by Ng. et al. (2011), using ammonium nitrate and ammonium sulfate. Air was sampled at the roof of the principal building at LAA (4.5 m above ground level) using a Teflon coated aluminum cyclone with a 2.5 µm

cut point at 5 LPM (model 1270; URG corporation, Chapel Hill NC). A multi-tube Nafion dryer (model PD-50T-12-MSS; Perma Pure LLC, Lakewood NJ) was installed right before the ACSM inlet to dry the aerosol. The data acquisition program was set for a time resolution of ~30 min. For the data analysis, a collection efficiency (CE) equal to 0.5 was used because this value was successfully used in previous campaigns in Mexico City using AMS technology. The uncertainties for the NR-PM₁ are estimated to be -30 to +10% (Salcedo et al., 2006; Aiken et al., 2009).

2.3 Other instruments

Along with the ACSM, other instruments were located on the roof of the LAA building from November 13, 2013 until March 31, 2014. Particulate matter total concentration was measured using two Tapered Element Oscillating Micro-Balances (TEOM, model 1400AB equipped with a Filter Dynamics Measurement System 8500C; Thermo Scientific, Franklin MA), with 2.5 and 1 µm cut point cyclones (VSCC; BGI, Butler NJ) for size selection. TEOM has a precision of $\pm 2.5 \ \mu g \ m^{-3}$ for 1-h averages. Additionally, en aethalometer (AE33 2008; Magee Scientific, Berkeley CA) with a selective inlet for PM2.5 (SCC1.829; BGI, Butler NJ) was used to detect black carbon (BC). The BC mass concentration was calculated from the change in optical attenuation at the 880 nm channel using a mass absorption cross section of 7.77 m^2g^{-1} . The precision of this measurement was < 10%.

In addition to the particle instruments, the following gases were measured: CO, SO₂, NO and NO₂, and O₃ (Serinus30, Serinus50, Serinus40, Serinus10, respectively; Ecotech Pty Ltd, Knoxfield Australia). Temperature (T) and relative humidity (RH) were measured using a sensor (083E; Met One Instruments, Grants Pass OR) located inside an aspirated radiation shield positioned 4 m above ground level on a meteorological tower. For wind speed (WS) and direction (WD) a lightweight three-cup anemometer and vane (010C and 020C; Met One Instruments, Grants Pass OR) were placed on top of the 10-m tower.

Figures S1 and S2, in the supplemental material, show the time series of the above meteorological variables and pollutants concentrations during the campaign.



Fig. 2. Time series and scatter plot of ACSM + BC and PM_1 mass concentrations at LAA. The vertical dotted arrow in panel (a) separates the acidic and neutral periods (see section 3.2). The solid line in panel (b) corresponds to the fitted line, and the dotted line to the 1:1 line.

2.4 General remarks

All concentrations are reported in local temperature and pressure conditions. The time used corresponds to Coordinated Universal Time minus 6 h (UTC-6). Before April 6 this time corresponded to local time. However, Daylight Savings Time started on April6; after that, local time corresponded to UTC minus 5 h (UTC-5). We used the built-in fitting routines included in Igor Pro 6.36 data analysis software (Wavewmetrics, 2014) for the statistical analysis. In all cases, we use Pearson's r (r) to describe the linear dependence between two variables.

3. Results and discussion

3.1 PM concentrations

The ACSM only detects the non-refractory material in PM₁; hence, assuming that BC was mainly in the submicrometric fraction (Schwarz et al., 2008), we added the ACSM total (sum of nitrate, sulfate, ammonium, chloride and organics) and the black carbon concentrations (ACSM + BC) in order to obtain a better estimate of the PM1 mass concentration. The result was compared to the PM1 TEOM measurements in Figure 2. There is a good linear correlation between ACSM + BC and PM_1 (r = 0.91), which is a sign of the soundness of the data. ACSM + BC concentration is $\sim 20\%$ lower than the PM₁ (the slope of the fitted line is 0.81), probably because of the refractory material (dust and metals) not accounted for in ACSM + BC. For example, in previous similar studies in Mexico City (MCMA-2003 and MILAGRO) it was estimated that the mineral material (which included metals and soil) accounted for approximately 7-8 % of the fine aerosol (Salcedo et al., 2006; Aiken et al., 2009). Experimental uncertainties of each instrument, as well as differences in inlets design and position, might also contribute to the difference observed.

Average concentrations of total NR-PM₁, BC, and PM concentrations measured in this study are shown in Table I. The average concentrations and mass fractions of the main NR-PM₁ components from this campaign, as well as from the MCMA-2003 (Salcedo et al., 2006) and MILAGRO (Querol et al., 2008; Aiken et al., 2009) campaigns are also included in the table, showing that all comparable values are on the same order of magnitude during the three campaigns. The differences observed might be explained by year-to-year variability and site location. The yearly variability of the PM concentrations in Mexico City is illustrated in Figure S3 in the supplemental material, where PM2.5 monthly concentrations measured in two sites of the Mexico City's Red Automática de Monitoreo Atmosférico (Air Quality Monitoring Network) from 2003 to 2015 are plotted (SEDE-MA, 2016b). Camarones (CAM; 19° 28' 06.18" N, 99° 10' 10.95" W) and UAM Iztapalapa (UIZ; 19° 21' 38.90" N, 99° 04' 25.96" W) sites were chosen because they are the nearest stations to LAA and CEN sites, respectively. The standard deviations of the concentrations for the same month in the same site were in the range of 10 to 20%, depending on the month; differences between the two sites were as large as 30% for the same monthly period. It is also important to note that PM concentrations shown in Table I for different campaigns were measured using different instruments such as TEOMs, optical particle counters (OPC, Grimm; Ainring, Germany), and Dusttrak Aerosol Monitors (TSI Inc.; Shoreview,

		LA	A	MILAGRO (T0) ¹		MCMA-2003 (CEN) ²		
		Nov. 13, 2013-Apr. 30, 2014		Mar. 1 - Ap	r. 4, 2006	Mar. 31 - May 4, 2003		
		$(\mu g m^{-3})$	%	$(\mu g m^{-3})$	%	$(\mu g m^{-3})$	%	
AMS	Organics	12.0	59.3	17.3	64.6	21.6	69.9	
NR-PM ₁	Sulfate	3.2	16.1	3.6	13.4	3.1	10.1	
	Nitrate	2.9	14.4	3.5	13.1	3.7	11.9	
	Ammonium	1.8	9.0	2.0	7.7	2.2	7.0	
	Chloride	0.2	1.2	0.4	1.5	0.3	1.0	
	ACSM total	20.2		26.8		30.9		
	BC	3.03 ^Ŧ		4.2*		3.4*		
	Soil			$1.7^{\$}$		2.1^{\pm}		
	Metals			1.0^{T}				
	PM _{2.5}	37.0 ^a		40.0 ^{b,3}		35.7 ^a , 40.0 ^c		
	PM ₁	27.8 ^a		33.0 ^{b,3}				

Table I. Average PM mass concentrations and percent composition from this study (LAA), and from MILAGRO and MCMA-2003 campaigns.

^T PM_{2.5}; * PM_{2.0}; [§] PM₁; ^a TEOM; ^b OPC; ^c Dusttrak; ¹ Aiken et al., 2009; ² Salcedo et al., 2006; ³ Querol et al., 2008.

MN). The variety of instruments used might also affect the comparisons among campaigns due to the inherent uncertainties of each of them. Finally, concentrations of the non-refractory material measured in the previous campaigns are also shown in Table I as a reference for the material that was not accounted for by the ACSM. Unfortunately, we do not have information on this fraction of the aerosol for the 2013-2014 campaign.

Data in Table I show that the average composition of NR-PM₁ during the campaign was similar to previous campaigns: i.e., the organic component is the largest one, followed by sulfate, nitrate, and ammonium. Chloride comprises only 1% of the mass. Figure 3 shows the time series of the concentrations of the main NR-PM₁ components and their mass fraction. In general, nitrate and ammonium showed a diurnal trend that was not as clear for the rest of the components. Nitrate concentration was higher in November and December compared to the rest of the campaign. Sulfate displayed periods of elevated concentrations that lasted few hours to several days; these periods were more common during the first half of the campaign. In general, chloride was present in very low concentrations; however, two large spikes were observed on December 25 and January

1, which were accompanied by high concentrations of sulfate, probably due to the Christmas and New Year's fireworks (Drewnick et al., 2006).

3.2 PM_1 acidity

In order to determine the acidity of the NR-PM₁, we calculated the predicted ammonium needed to completely neutralize the sulfate, nitrate and chloride present in the aerosol, using the following equation:

$$NH_{4_pred} = 18\left(\frac{2[SO_{4}^{2-}]}{96} + \frac{[NO_{3}^{-}]}{62} + \frac{[Cl^{-}]}{35.5}\right) \quad (1)$$

where $[SO_4^{2-}]$, $[NO_3^{-}]$, and $[CI^{-}]$ are the concentrations of sulfate, nitrate, and chloride in $\mu g m^{-3}$. This equation assumes that the inorganic anions are in the form of neutralized ammonium salts: NH₄NO₃, (NH₄)₂SO₄, NH₄Cl. We also calculated the missing ammonium for neutrality using the following equation:

$$NH_4 miss = NH_4 pred - NH_4 meas$$
 (2)

where NH_4 meas is the measured ammonium concentration in $\mu g m^{-3}$. The time series of NH_4 miss



Fig. 3. Time series of the main NR-PM₁ components and their mass fraction at LAA. The vertical dotted arrow separates the acidic and neutral periods (see section 3.2).

is presented in Figure 4a, where it is evident that NH_4 _miss was persistently higher at the beginning of the campaign than at the end (see also Table II and Fig. 5). In fact, when plotting NH_4 vs NH_4 _pred, the data points before January 2, 2014 (Fig. 4b) lie below the 1:1 line, indicating that there is not enough ammonium to neutralize the sulfate, nitrate and chloride; i.e. the aerosol is acidic. In fact, we recalculated the predicted ammonium assuming that sulfate is in the form of ammonium bisulfate (NH_4HSO_4):

$$NH_4_pred_bs = 18\left(\frac{[SO_4^{2-}]}{96} + \frac{[NO_3^{-}]}{62} + \frac{[C1^{-}]}{35.5}\right)(3)$$

When NH₄_pred_bs is plotted vs. NH₄_meas (Fig. 4c), the fitted line has a slope closer to 1. After January 2, 2014 (Fig. 4d), the aerosol can be considered neutral.

The results for the second period of the campaign are consistent with observations during MCMA-03 and MILAGRO campaigns, when the aerosol was also found to be neutral in March and April 2003 and 2006 (except for one week in 2003) (Salcedo et al., 2006; Aiken et al., 2009). However, this is the first report of a persistent acidity in the Mexico City aerosol. It would be important to repeat this kind of long-term campaigns for several years in order to determine if this is a yearly trend.

Table II includes the average concentrations of ACSM components during the acidic and neutral periods, as described in the previous paragraph. During the acidic period, sulfate and nitrate concentrations were larger than during the neutral period, while ammonium concentrations were similar, which explains the difference in acidity observed.

In order to look for the conditions that might be responsible for the observed change in acidity, the Kruskal-Wallis (KW) test was used (with a confidence level of 99%) to determine if there were statistically significant differences in the parameters measured during the two periods. We also draw box-and-whisker plots of the same parameters for an easier visualization of the differences. Figure 5 shows the plots for the parameters that might be more related with aerosol acidity, and Figure S4 for the



Fig. 4. (a) Time series of NH4_miss (Eq. 2). (b) and (c) NH4_meas vs. NH4_pred, and vs. NH4_pred_bs (Eqs. 1 and 3) during the acidic period. (d) NH4_meas vs. NH4_pred (Eq. 1) during the neutral period. The solid lines in panels b-d correspond to the fitted lines, and the dotted lines to 1:1 lines.

	Acidic period Nov. 13, 2013-Jan. 1, 2014					Neutral period Jan. 2-Apr. 30, 2014				
	(µg m ⁻³)					(µg m ⁻³)				
	Average	SD^1	Min.	Max.	%	Average	SD^1	Min.	Max.	%
NH ₄ _miss	0.81	0.92	-1.2	6.8		0.13	0.35	-1.4	4.0	
Organics	12.4	7.3	0.78	64.7	55.7	11.8	6.3	0.25	70.4	61.6
Sulfate	4.1	3.9	_	26.3	18.3	2.8	2.0	_	16.5	14.7
Nitrate	3.6	4.1	0.03	27.4	16.1	2.6	2.8	0.03	18.2	13.4
Ammonium	1.9	1.5	_	9.5	8.6	1.8	1.3	_	10.7	9.3
Chloride ACSM total	0.3 22.2	0.54 13.5	1.5	5.5 103.3	1.4	0.2 19.1	0.3 10.0	1.1	4.2 87.9	1.1

Table II. Summary of the composition of $NR-PM_1$ for the acidic and neutral periods at LAA. Minimum values without a value indicate signals within the noise level.

¹ Standard deviation.

rest. The KW test indicated that all parameters were statistically different between periods, except for ammonium, organics, and wind speed. These results confirm the discussion regarding concentrations of NH₄_miss, nitrate, sulfate and ammonium discussed above. SO₂ and NO_x concentrations, precursors of sulfate and nitrate respectively, were higher during the acidic period. In addition, the RH was higher, and the T lower; both of which conditions can be related

to higher concentrations of sulfate and nitrate due to faster oxidation rates, and the HNO₃ gas-particle equilibrium shifting to the condensed phase (Seinfeld and Pandis, 2012). Finally, the wind roses shown in Figure 6 reveal more frequent, and faster winds from the NW affecting the LAA during the acidic period. These winds might also affect the amount of sulfate observed because they come from the direction of the Tula industrial complex (see Fig. 1), which is one of



Fig. 5. Box-and-whisker plots for several parameters measured during the acidic (grey) and neutral (black) periods. Whiskers correspond to 5 and 95 percentiles; vertical lines to 25, 50, and 75 percentiles; and circles to the average value. KW tests indicated that all parameters shown here were statistically different between periods, except for ammonium.



Fig. 6. Wind roses during the acidic and neutral periods. The radial axes in the wind rose indicate frequency of winds.

the most important SO_2 sources in Mexico City (de Foy et al., 2009). In conclusion, the higher acidity observed at the beginning of the campaign might be related to higher SO_2 and NO_x concentrations, higher RH, lower T, and more frequent winds from the NW. In fact, high relative humidity and northerly winds were also observed during the week in 2003 when acidic aerosols were detected during MCMA-2003 (Salcedo et al., 2006).

In Figure 7 we plotted NH₄_miss vs. sulfate and nitrate. While NH₄_miss shows a good correlation with sulfate (r = 0.9093), it does not correlate at all with nitrate, which can be explained by the complex



Fig. 7. NH4_miss (Eq. 2) vs. sulfate and nitrate during the acidic period. The black line in the left panes corresponds to the linear fit.

phase and acid-base equilibriums of the system $(H_2SO_4 + HNO_3 + NH_3 + H_2O)$ (Seinfeld and Pandis, 2012). Sulfuric acid is non-volatile and exists mainly in the condensed phase (solid or liquid); on the other hand, nitric acid and ammonia partition between the gas and the condensed phases, depending on the RH, temperature, and the total concentration of all the species in the system. In general, under ammonia-rich conditions, NH₃ will neutralize H₂SO₄ and HNO₃ to form sulfate, nitrate, and ammonium, which can be aqueous or forming salts depending on the RH. However, in the case of insufficient NH₃ the sulfuric acid will tend to be partially neutralized (as bisulfate) and the nitric acid will be forced to the gaseous phase. During the MILAGRO campaign, Fountoukis et al. (2009) found that the atmosphere in the MCMA was unusually ammonia-rich, which was used to explain the neutrality of the aerosol back then. In contrast, Figure 7 suggests that, during November and December 2013, the ammonia present in the atmosphere was not enough to neutralize the nitric and sulfuric acids. This suggestion is consistent with results from Cady-Pereira et al. (2017), who calculated NH₃ concentrations over Mexico City during 2013-2015 using the Tropospheric Emission Spectrometer (TES)

instrument on the NASA AURA satellite, and found a seasonal trend in NH₃ concentrations, with a minimum during September-November and a maximum in March-May. Cady-Pereira et al. (2017) results not only shed light on the reason for the aerosol acidity observed during 2013, but indicate a probable yearly trend. Again, it would be important to perform more campaigns to verify this hypothesis.

The acidity of the aerosol has wide implications that span from effects to human and ecosystems health, to measures to control particle concentrations through emissions regulation. Also, the pH of aerosols can change the chemistry occurring in the atmosphere through heterogenous acid-catalyzed reactions, and increase solubility of material associated with mineral dust (Weber et al., 2016). Hence, it is important to better understand the variability of the PM acidity, as well as the factors that determine it, which can only be done with more long-term studies such as this one.

3.3 NR-PM₁ diurnal cycles

Average diurnal plots of the main NR-PM₁ components in 2013-2014, 2006 (Aiken et al., 2009), and 2003 (Salcedo et al., 2006) are shown in Figure 8.



Fig. 8. Average diurnal cycles of the NR-PM₁ components at LAA (acidic and neutral periods), and during MILAGRO at T0 (Aiken et al., 2009), and MCMA-2003 at CEN (Salcedo et al., 2006).

The plots show the hourly average concentrations of all the data available for each period. Figures S5 and S6 show the diurnal box-and-whisker plots for the diurnal concentrations during the acidic and neutral periods, respectively. In general, as for the average concentrations discussed in section 3.1, the behavior of all the components were similar during the three campaigns. Nitrate shows a maximum concentration in the morning, which corresponds to the photochemical activity, and it is closely followed by ammonium. The organic fraction also shows a maximum at midday, but the raise starts earlier in the morning due to the primary emissions by vehicles; i.e., its diurnal cycle is a combination of primary aerosols emitted during the early rush hour, and secondary aerosols formed during the morning. Chloride maximum concentrations occur in the early morning due to its relatively high volatility. Finally, the diurnal cycle of sulfate is determined by its regional nature. The diurnal cycles during acidic and neutral periods are qualitatively similar; however, the absolute values are different, especially for nitrate and sulfate, as discussed in section 3.2.

The sulfate diurnal cycle is compared to the NH_4 _miss cycle in Figure 9, where it is evident that there is a close relation between the acidity of the aerosol and the sulfate concentration during the



Fig. 9. Average diurnal cycles of the NH4_miss and sulfate concentrations for the acidic and neutral periods.

acidic period. Such relation is not observed during the neutral period, which is consistent with the discussion in section 5.2.

4. Conclusions

Mass concentration and composition of the NR-PM₁ was measured with a 30-min time resolution in a site north of Mexico City for a period of almost six months. Average PM₁ concentrations and composition in 2013-2014 were in the same order of magnitude as observed in previous campaigns during the spring of 2003 and 2006. The differences observed might be explained by yearly variability and the different location of sites. However, in November and December 2013 the aerosol showed a persistent acidic behavior due to a larger concentration of sulfate and nitrate with respect to ammonium. This observation is consistent with higher RH and lower T during these two months, which promote faster oxidation rates and the HNO₃ gas-particle equilibrium shifting to the condensed phase. More frequent winds from the NW, where Tula industrial complex is located, might also affect the acidity of the aerosol, because Tula represents one of the main SO₂ sources in Mexico City. Finally, a seasonal reduction in the concentration of gas phase ammonia during November and December probably inhibited the neutralization processes of the aerosol.

This study represents the longest continuous campaign measuring aerosol composition in Mexico City with a high time resolution. The results suggest a seasonal variability in aerosol acidity that had not been reported before, which could have important implications on human and ecosystems health, as well as on the chemistry occurring in the atmosphere. It would be necessary to perform more long-term campaigns in order to determine if this is a yearly trend. Furthermore, this study exposes the little knowledge that exists regarding chemical processes occurring under different atmospheric conditions in Mexico City, and indicates the importance of longterm studies.

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Supplemental material

Fig. S1. Time series of the relative humidity (% RH), temperature (T), and wind speed (WS) at LAA. The vertical dotted arrow separates the acidic and neutral periods (see section 3.2).



Fig. S2. Time series of criteria pollutants concentrations at LAA. The vertical dotted arrow separates the acidic and neutral periods (see section 3.2).



Fig. S3. Time series of the average PM_{2.5} monthly concentration, at two sites of Mexico City's Red Automática de Monitoreo Atmosférico (Air Quality Monitoring Network) (SEDEMA, 2016). UIZ is the closest site in the network to CEN; CAM is the closest to T0 and LAA.



Fig. S4. Box-and-whisker plots for several parameters measured during the acidic (grey) and neutral (black) periods. Whiskers correspond to 5 and 95 percentiles; vertical lines to 25, 50, and 75 percentiles; and circles to the average value. KW tests indicated that all parameters shown here were statistically different between periods, except for organics, and wind speed.



Fig. S5. Diurnal box-and-whisker plots of the NR-PM₁ main species mass concentration, during the acidic period. Whiskers correspond to 5 and 95 percentiles; vertical lines to 25, 50, and 75 percentiles; and circles to the average value.



Fig. S6. Diurnal box-and-whisker plots of the NR-PM₁ main species mass concentration, during the neutral period. Whiskers correspond to 5 and 95 percentiles; vertical lines to 25, 50, and 75 percentiles; and circles to the average value.